

1 **Stoichiometric constraints modulate the effects of temperature and nutrients on biomass**
2 **distribution and community stability**

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15 **Author contributions**

16 A.S., B.H., and J.M.M. conceived the study. B.H. and A.S. developed and analysed the models.

17 A.S. wrote the first draft of the manuscript. All authors contributed substantially to revisions.

18

19 **Running title:** Stoichiometry alters thermal effects in food webs

20 **Key-words:** temperature, stoichiometry, temporal variability, paradox of enrichment, trophic
21 interactions, nutrient quota, consumer-resource dynamics, biomass structure.

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23 **Legal and ethical requirements**

24 The authors declare that they have no conflict of interest.

25

26 **Abstract**

27 Temperature and nutrients are two of the most important drivers of global change. Both can
28 modify the elemental composition (i.e. stoichiometry) of primary producers and consumers. Yet
29 their combined effect on the stoichiometry, dynamics, and stability of ecological communities
30 remains largely unexplored. To fill this gap, we extended the Rosenzweig-MacArthur consumer-
31 resource model by including thermal dependencies, nutrient dynamics, and stoichiometric
32 constraints on both the primary producer and the consumer. We found that stoichiometric
33 constraints dampen the paradox of enrichment and increased persistence at high nutrient levels.
34 Nevertheless, they also reduced consumer persistence at extreme temperatures. Finally, we also
35 found that stoichiometric constraints can strongly influence biomass distribution across trophic
36 levels by modulating consumer assimilation efficiency and resource growth rates along the
37 environmental gradients. Our findings highlight the importance of accounting for stoichiometric
38 constraints as they can mediate the temperature and nutrient impact on the dynamics and
39 functioning of ecological communities.

40
41 **Key words:** temperature, stoichiometry, temporal variability, paradox of enrichment, trophic
42 interactions, nutrient quota, consumer-resource dynamics, biomass structure.

43 **Introduction**

44 Temperature and nutrients regulate many biological processes, including species geographical
45 distribution, primary production, species interactions, and energy and material fluxes (Falkowski
46 *et al.* 1998; Enquist *et al.* 1999; Elser *et al.* 2007; Thomas *et al.* 2017). They are at the core of
47 several ecological theories. While temperature is a fundamental component of metabolic scaling
48 theory (Brown *et al.* 2004), nutrients are at the core of resource competition theory (Tilman 1982)
49 and ecological stoichiometry (i.e. the element composition of organisms) theory (Sturner & Elser
50 2002). Cross *et al.* (2015) suggested that a better understanding of the interactions between
51 temperature and nutrients is crucial for developing realistic predictions about ecological
52 responses to multiple drivers of global change, including climate warming and elevated nutrient
53 supply. Nutrients can modulate the effects of warming on communities directly by altering
54 primary production, and/or indirectly by changing the elemental composition of primary
55 producers. Conversely, thermal effects on trophic interaction strengths (i.e. the *per capita* effect
56 of predators on prey population densities) and on consumer energetic efficiencies (i.e. ingestion
57 relative to metabolic demand) depend on both the quantity and quality of their resources. While
58 Cross *et al.* (2015) provided a road map on how to investigate the combined effects of
59 temperature and nutrient on ecological processes, we still lack an integrative theory to better
60 understand how the links between stoichiometry, nutrient enrichment, and temperature influence
61 the dynamics and stability of multispecies communities. Such a theory will allow us to
62 understand how and when stoichiometric variation modulates the consequences of single and
63 combined components of global change on trophic interactions, community dynamics, and
64 ecosystem functioning.

65
66
67 Predicting the effects of global warming and nutrient changes on ecosystems is challenging as
68 species are embedded within communities of multiple interacting species (Petchey *et al.* 1999;

69 Tylianakis *et al.* 2008; Montoya & Raffaelli 2010; Gilbert *et al.* 2014). Increased resource
70 availability (hereafter: enrichment) and warming can jointly affect food-web stability and
71 structure by modifying the strength of trophic interactions (O'Connor *et al.* 2009; Binzer *et al.*
72 2012; Kratina *et al.* 2012; Sentis *et al.* 2014; Binzer *et al.* 2016). Enrichment typically increases
73 energy flux from resources to higher trophic levels which often leads to the well-known paradox
74 of enrichment where the amplitude of population fluctuations increase with nutrients, leading to
75 extinctions at high nutrient concentrations (Rosenzweig 1971; Rip & McCann 2011; Gilbert *et al.*
76 2014). Nevertheless, most consumer species become less efficient at processing matter and
77 energy at warmer temperatures as their metabolic rates often increase faster with temperature
78 than their feeding rates (Vucic-Pestic *et al.* 2011; Fussmann *et al.* 2014; Iles 2014). This
79 reduction of energetic efficiency lessens energy flow between trophic levels and hence stabilizes
80 food-web dynamics by reducing population fluctuations (Rip & McCann 2011; Binzer *et al.*
81 2012; Gilbert *et al.* 2014). As a result, mild warming may alleviate the paradox of enrichment by
82 decreasing consumer energetic efficiency (Binzer *et al.* 2012; Sentis *et al.* 2017).

83
84 The theoretical expectations and results described above have already improved our ability to
85 understand and predict the effects of temperature and enrichment on food webs (Boit *et al.* 2012;
86 Tabi *et al.* 2019). However, most previous studies using metabolic scaling theory assumed that
87 nutrient enrichment lead to an increase in resource carrying capacity without influencing resource
88 elemental composition (Vasseur & McCann 2005; Binzer *et al.* 2012; Gilbert *et al.* 2014; Binzer
89 *et al.* 2016; Sentis *et al.* 2017). Yet nutrient enrichment effects are more complex. The elemental
90 composition of primary producers is likely to be altered, in response to the supplies of energy and
91 materials relative to their growth and nutrient intake rates (Rastetter *et al.* 1997;
92 Robert W. Sterner *et al.* 1997; Finkel *et al.* 2009) Sterner *et al.* 1997). This, in turn, can affect the
93 dynamics of the producer population and the herbivores feeding on it. For instance, previous

94 modelling studies showed that introducing stoichiometric heterogeneity in predator-prey
95 population dynamic models can dampen the negative effect of nutrient enrichment on system
96 persistence by reducing population biomass fluctuations (Andersen 1997; Loladze *et al.* 2000;
97 Andersen *et al.* 2004; Elser *et al.* 2012). More generally, the stoichiometric flexibility of primary
98 producers, in particular the flexibility in carbon to nutrient ratios (e.g. C:N or C:P), has important
99 implications for animal feeding behaviour (White 1993), consumer population stability (White
100 1993; Sterner & Hessen 1994; Hessen *et al.* 2002), community structure (Andersen 1997), and
101 ecosystem processes such as biogeochemical cycling (Andersen 1997; Hessen *et al.* 2004).

102
103 Previous theoretical and empirical studies reported that stoichiometric variations can have a
104 strong influence on the stability of consumer-resource interactions (Andersen 1997; Andersen *et*
105 *al.* 2004; Diehl *et al.* 2005; Elser *et al.* 2012). For instance, populations of crustacean *Daphnia*
106 feeding on low quality (i.e. low nutrient: carbon ratio) algae cannot persist even when resource
107 quantity is not a limiting factor (Elser *et al.* 2007). Consumer extinction is explained by the fact
108 that the consumer assimilation efficiency is, for most organisms, a function of resource quality
109 (Elser *et al.* 2000). When resource quality is low, the consumers assimilate only few nutrients
110 relative to the biomass they ingest, which limits their growth and reproduction (Elser *et al.* 2000;
111 Elser *et al.* 2012). Temporal variations in resource quality can stabilize the system by weakening
112 interaction strength and dampening population fluctuations (Andersen *et al.* 2004; Diehl *et al.*
113 2005)but see(Loladze *et al.* 2000; Elser *et al.* 2012). However, it remains unclear whether and
114 how temporal variations in the elemental composition of primary producers and consumers can
115 modulate the effects of temperature and nutrients on important community features such as
116 stability and biomass distribution across trophic levels. Previous studies indicated that the spatial
117 and temporal intraspecific variations in the elemental composition of primary producers are
118 expected to increase in response to global change drivers such as temperature, CO₂, and nutrient

119 availability (Bezemer & Jones 1998; Woods *et al.* 2003; Finkel *et al.* 2009). This increased
120 variation can be of importance for both primary producer and consumer populations as the
121 growth rate of primary producers is well known to depend on their elemental composition (Droop
122 1974) as is the assimilation efficiency of the consumers (Sterner & Elser 2002).

123
124 Altogether, previous studies indicated that both temperature and stoichiometric variations can
125 have important effects on species interactions and community dynamics (Andersen *et al.* 2004;
126 Diehl *et al.* 2005; Fussmann *et al.* 2014; Binzer *et al.* 2016; Sentis *et al.* 2017). However, the
127 effects of temperature and nutrient stoichiometry on food web dynamics and stability have only
128 been studied in isolation. Recent theory by Uszko *et al.* (2017) showed that considering nutrient
129 dynamics can help to better understand the influence of temperature on consumer-resource
130 population dynamics and resource carrying capacity. Nevertheless, they considered that the
131 elemental composition of both the resource and the consumer are constant and independent of
132 temperature and nutrient dynamics. This contrasts with the empirical observation that resource
133 elemental composition is flexible and can vary with both temperature and nutrient dynamics
134 (Droop 1974; Elser *et al.* 2000; Woods *et al.* 2003). Here we thus focused on the combined
135 effects of temperature and nutrients on the stoichiometry of primary producers and how this
136 affects community stability and biomass distribution across trophic levels in a consumer-resource
137 system. Understanding the determinants of stability and biomass distribution has been at the core
138 of ecology for a long time (Elton (1927), Lindeman (1942)). Recent theory aims at explaining
139 empirical observations of trophic pyramids (i.e. population biomass decreases with trophic
140 levels), inverted trophic pyramids (i.e. population biomass increases with trophic levels), trophic
141 cascades and the link between biomass distribution and stability (McCauley *et al.* 2018; Barbier
142 & Loreau 2019).

143

144 Here, we used the Rosenzweig-MacArthur model as a baseline non-stoichiometric model because
145 this model is one of the most studied models used to investigate the effects of temperature and
146 nutrient enrichment on community dynamics (Vasseur & McCann 2005; Binzer *et al.* 2012;
147 Fussmann *et al.* 2014; Sentis *et al.* 2017). Inspired by previous temperature-independent
148 stoichiometric consumer-resource models (Andersen 1997; Andersen *et al.* 2004; Diehl *et al.*
149 2005), we then extended the Rosenzweig-MacArthur model to account for nutrient dynamics as
150 well as for the simultaneous dependence of community dynamics on temperature and flexible
151 resource stoichiometry. Our objective here was not to develop a complex and very realistic
152 stoichiometric model that would include additional important abiotic and biotic features such as
153 light intensity (Diehl 2007) or compensatory feeding (Cruz-Rivera & Hay 2000). Instead, we
154 aimed at introducing two fundamental stoichiometric features (i.e. stoichiometric flexibility and
155 stoichiometric imbalance) and investigate how these stoichiometric considerations can change
156 predictions of the Rosenzweig-MacArthur model. We thus used our extended Rosenzweig-
157 MacArthur model to predict the effects of warming and nutrient enrichment on population
158 dynamics and biomass distribution across trophic levels and compared these predictions with the
159 predictions of the nonstoichiometric Rosenzweig-MacArthur model. We particularly addressed
160 two questions: (i) How do stoichiometric constraints modulate the effects of enrichment and
161 warming on community stability and persistence? and (ii) How do stoichiometric constraints
162 modulate the effects of enrichment and warming on biomass distribution across multiple trophic
163 levels?

164 **Methods: Population dynamic models**

165 **The Rosenzweig-MacArthur (RM) model.**

166 Rates of change of the consumer and resource biomass densities \dot{C} and \dot{R} depend on their
167 respective biomass densities C and R ($\text{g}\cdot\text{m}^{-3}$):

168
$$\dot{R} = r\left(1 - \frac{R}{K}\right)R - \frac{aR}{1+ahR}C \quad (1)$$

169
$$\dot{C} = \left(e\frac{aR}{1+ahR} - m\right)C \quad (2)$$

170 The population growth rate of the resource is given by the logistic equation where r is the
171 resource maximum growth rate and K is the resource carrying capacity. The population growth
172 rate of the consumer is equal to its feeding rate multiplied by its assimilation efficiency e (i.e. the
173 fraction of resource biomass converted into consumer biomass) minus a loss term associated to
174 metabolic losses m . The feeding rate of the consumer C depends on the density of its resource R
175 and follows a Holling type II functional response, with consumer-resource attack rate a and
176 handling time h . Our choice for a type II functional response is motivated by a meta-analysis
177 reporting that most consumers feed on their prey following a saturating type II functional
178 response rather than a linear type I or a sigmoidal type III functional response (Jeschke *et al.*
179 2004).

180
181 In the RM model, consumer and resource population growth rates are only limited by nutrient or
182 resource density. Nutrient enrichment is assumed to increase resource carrying capacity, which
183 often leads to the well-known paradox of enrichment where populations fluctuates up to
184 extinctions (Rosenzweig 1971). Nevertheless, this model neither considers nutrient dynamics nor
185 temporal variations of resource stoichiometry and their consequences on population dynamics.
186 To circumvent these limitations of the RM model, we extended it to better consider nutrient

187 dynamics, resource stoichiometry and the way they can affect resource and consumer population
188 dynamics.

189

190 **The Stoichiometric Rosenzweig-MacArthur (SRM) model.**

191 We derived a stoichiometric extension of the Rosenzweig-MacArthur consumer–resource model
192 with additional stoichiometric and temperature dependencies of several biological rates. We
193 considered two stoichiometric constraints: one on the resource population growth rate, and the
194 other on the consumer assimilation efficiency (see below for more details). These stoichiometric
195 constraints have been observed for several consumer-resource pairs suggesting that they are core
196 components of species growth and interactions (Sterner & Elser 2002).

197

198 *Stoichiometric constraint on the resource population growth rate*

199 Inspired by previous stoichiometric models (Andersen 1997; Loladze *et al.* 2000; Andersen *et al.*
200 2004; Diehl *et al.* 2005), we extended the RM model by considering explicit nutrient dynamics
201 and nutrient effects on resource population growth rate. The system is assumed to be closed for
202 nutrients. Thus, nutrient supply originates exclusively from biomass excretion and
203 remineralization. The total amount of nutrients in the system (N_{tot}) is then a measure of nutrient
204 enrichment. In contrast to the very high plasticity in C:N or C:P exhibited by autotrophs,
205 heterotrophs regulate elemental composition within narrower bounds, even when consuming food
206 with large variation in elemental composition (Andersen & Hessen 1991; Sterner & Hessen 1994;
207 Andersen 1997; Elser *et al.* 2000). In other words, the elemental homeostasis is much stronger for
208 consumers compared to primary producers. We thus assumed the nutrient quota (i.e. the nutrient
209 to carbon ratio) of the consumer Q_C to be conserved whereas the one of the resource Q_R is
210 flexible over time with the only constraint that $Q_R > 0$. As in the RM model, rates of change of
211 the consumer and resource biomass densities \dot{C} and \dot{R} depend on their respective carbon biomass

212 densities C and R ($\text{gC}\cdot\text{m}^{-3}$), except that the resource population growth rate follows the Droop
213 equation (Droop 1974) given by $r(1-Q_{\min}/Q_R)R$ and is now limited by Q_R relative to the
214 minimum nutrient quota Q_{\min} :

$$215 \quad N_{tot} = Q_R R + Q_C C \quad (3)$$

$$216 \quad \dot{R} = r\left(1 - \frac{Q_{\min}}{Q_R}\right)R - \frac{aR}{1+ahR}C \quad (4)$$

$$217 \quad \dot{C} = \left(e \frac{aR}{1+ahR} - m\right)C \quad (5)$$

218 From the nutrient conservation equation (eqn 3) we obtain that $Q_R = \frac{N_{tot}-Q_C C}{R}$. The intuitive
219 interpretation is that the resource nutrient quota Q_R changes instantaneously with the density of
220 the resource population R and with the density of the nutrient stored in the consumer biomass
221 $Q_C C$, to maintain nutrient balance (see Text S1 for details).

222

223 *Stoichiometric constraint on the consumer population growth rate*

224 In the RM model, the growth rate of the consumer population only depends on resource density.
225 We relaxed this assumption by making the population growth rate of the consumer dependent on
226 both resource quality (i.e. nutrient quota) and quantity (i.e. biomass density). In the SRM model,
227 consumer production is also limited by resource quality as the consumer assimilation efficiency e
228 is a saturating function of resource nutrient quota Q_R :

$$229 \quad e(Q_R) = e_{max} \frac{Q_R}{Q_R + Q_C} \quad (6)$$

230 The intuitive interpretation of eqn. 6 is that resource quality is not a limiting factor for consumer
231 growth as long as the nutrient content of the resource is superior to the nutrient content of the
232 consumer (i.e. $Q_R > Q_C$). In other words, $e(Q_R)$ is at its maximum for $Q_R > Q_C$ and proportional to
233 Q_R for $Q_R < Q_C$. By replacing e by $e(Q_R)$ in eqn. 5, we obtain the SRM model.

234

235 **Temperature dependence of model parameters**

236 To investigate the effect of temperature and stoichiometric constraints on consumer-resource
237 dynamics, we next extended the RM and SRM models described above by adding thermal
238 dependencies of the parameters. Following Uszko et al. (2017), we assumed that the total amount
239 of nutrient N_{tot} , the maximum food conversion efficiency e_{max} , and fixed stoichiometric traits
240 (Q_C) are independent of temperature, as there is no evidence of systematic temperature
241 dependence for any of them (Peters 1983; Ahlgren 1987; Borer *et al.* 2013; Yvon-Durocher *et al.*
242 2015). Rate of maintenance respiration and natural background mortality m typically increases
243 exponentially with temperature (Fig. S1a and b). We thus used the Arrhenius equation to describe
244 the effect of temperature T (in Kelvin) on m :

$$245 \quad m(T) = m_0 e^{\frac{-E_m}{kT}} \quad (7)$$

246 where m_0 is a parameter-specific constant calculated at temperature of 0°C (= 273.15 K). The
247 temperature dependence is characterized by the respective activation energy E_m (eV) and the
248 Boltzmann constant $k=8.62 \times 10^{-5}$ eVK⁻¹. As the temperature dependencies of resource intrinsic
249 growth rate r and functional response parameters (a , $1/h$) are often unimodal rather than
250 exponential (Englund *et al.* 2011; Rall *et al.* 2012; Sentis *et al.* 2012; Thomas *et al.* 2012), we
251 used Gaussian functions for r and a and an inverted Gaussian function for h :

$$252 \quad g(T) = g_0 e^{\pm \frac{(T-T_{\text{opt}})^2}{2s^2}} \quad (8)$$

253
254 where T_{opt} is the temperature at which the rate g reaches its minimum or maximum, s is the
255 function width and g_0 is a parameter-specific constant calculated at T_{opt} . The minus-sign
256 corresponds to Gaussian functions and the plus-sign to inverted Gaussian functions.

257

258 **Model parameterisation and simulations**

259 To parameterise the models we assumed the resource and consumer species to be a unicellular
260 freshwater algae and a *Daphnia* grazer, respectively. The choice for this system was motivated by
261 the good characterization of both the stoichiometric parameters and thermal dependencies for this
262 system (Andersen 1997; Uszko *et al.* 2017). Uszko *et al.* (2017) recently estimated the thermal
263 dependencies for biological rates of the green algae *Monoraphidium minutum* and the grazer
264 *Daphnia hyalina*. We thus used their estimates of stoichiometric parameters and thermal
265 dependencies (See Table S1 and Fig. S1 for further details).

266
267 To investigate the individual and combined effects of enrichment, warming, and stoichiometric
268 constraints, we varied temperature (401 values ranging from 0 to 40°C by 0.1°C) and total
269 amount of nutrients (parameter N_{tot} in eqn. 10; 60 values ranging from 0.001 to 0.06 gP.m⁻³ by
270 0.001 gP.m⁻³, overlapping with reported mean phosphorus concentration in European peri-alpine
271 lakes (Anneville *et al.* 2005)). For the RM model, we used the minimum nutrient quota to convert
272 nutrients into resource (i.e. $K = N_{\text{tot}}/Q_{\text{min}}$). This implies that carrying capacity is independent of
273 temperature which is expected for closed, nutrient-limited systems (Uszko *et al.* 2017) although
274 more experimental evidence are needed to verify this assumption. We then simulated the
275 consumer-resource dynamics for 1000 days to enable the system to reach an attractor (either an
276 equilibrium point or a limit cycle) before we assessed the final state. Therefore, for each model,
277 we simulated 24060 combinations of environmental conditions (401 temperatures by 60 nutrient
278 concentrations). Initial biomass density of each species was set to 0.98 times its equilibrium
279 density in the two-species system (calculated by solving for the two-species equilibrium, using
280 either eqns 1-2 for model RM or eqns 3-5 for model SRM). The value of 0.98 was chosen to be
281 (1) close enough to equilibria to avoid extinctions caused solely by transient dynamics and (2) not
282 exactly the equilibrium value to probe the stability of the equilibrium. Any population falling
283 below the extinction threshold of 10⁻⁹ g.m⁻³ during the simulation was deemed extinct and its

284 biomass set to zero to exclude ecologically unrealistic low biomass densities. For each model, we
285 calculated system persistence as the percentage of simulations with the two species remaining
286 extant at the end of the simulations. Population dynamics were simulated with R version 3.4.3 (R
287 Development Core Team 2017) using the “deSolve” package (Soetaert *et al.* 2012) with an
288 absolute error tolerance of 10^{-10} and a relative error tolerance of 10^{-6} .

289

290 **Results**

291 *Stability: population fluctuations and persistence*

292 Stoichiometric constraints dampened the paradox of enrichment, reducing fluctuations at high
293 nutrient levels and hence increasing persistence. However, stoichiometric constraints also
294 reduced the persistence of the consumer at low and high temperatures. As a result, the overall
295 effect of stoichiometric constraints on stability depends on its relative influence on population
296 fluctuations versus consumer persistence. In the two following paragraphs, we explain in more
297 detail these results and highlight key differences between the outcomes from RM and SRM
298 models.

299

300 The RM model predicts that increasing nutrient concentration is strongly destabilizing: the
301 system shifts from a stable equilibrium point to limit cycles (i.e. the system crosses a Hopf
302 bifurcation). This agrees with the paradox of enrichment. As population biomass fluctuations (i.e.
303 cycle amplitude) increase with nutrient concentration, minimal population densities are very low
304 at high nutrient concentrations leading to the extinction of both the consumer and resource once
305 the extinction threshold is crossed (Fig. 1). In the range of temperatures where the consumer
306 persists, warming does not have a strong influence on the nutrient concentration at which the
307 system shifts from the stable equilibrium point to limit cycles, although this qualitative shift is

308 absent at very high temperatures (i.e. 32°C) when the consumer is close to extinction. Warming
309 decreases fluctuation amplitude and thus dampens extinctions driven by the paradox of
310 enrichment, which results in warming enhancing the persistence of the consumer-resource system
311 at high nutrient concentrations. However, very warm and cold temperatures cause the extinction
312 of the consumer (see below for the mechanisms underlying extinctions), releasing resources from
313 top-down control. Overall, we found that, without considering the extinction threshold of 10^{-9}
314 g.m^{-3} (see Model parametrisation and simulations), both the consumer and the resource can
315 persist in 74% of the temperature-nutrient concentration scenarios (i.e. black + orange areas in
316 Fig 1C). Nevertheless, when considering the extinction threshold, they persist in only 21% of the
317 temperature-nutrient scenarios (i.e. black area in Fig. 1c) indicating that extinctions driven by
318 population fluctuations are highly prevalent in the RM model.

319
320 In contrast, the SRM model shows that increasing nutrient concentrations causes fewer
321 fluctuations than those observed for the RM model (Fig. 1). This is because: (1) more nutrients
322 are needed to shift the system from a stable equilibrium point to limit cycles—the system can
323 indeed persist without fluctuations up to 0.02 gP.m^{-3} whereas it was only up to 0.0005 gP.m^{-3} in
324 the RM model—and (2) when the system fluctuates, the amplitude of the fluctuations is smaller
325 in the SRM than in the RM model. As a result, stoichiometric constraints dampen the amplitude
326 of population fluctuations (i.e. the paradox of enrichment) and hence increase system persistence
327 at high nutrient levels. While the qualitative effect of temperature is similar to that observed in
328 the RM model, the thermal thresholds for consumer persistence are reduced at low and high
329 temperatures in the SRM predictions. Moreover, thermal thresholds remain almost constant along
330 the nutrient gradient in the RM model, whereas in the SRM model they depend on nutrient
331 concentration, with a smaller thermal range at low nutrient levels compared to high nutrient
332 levels (Fig. 1). The consumer is thus more likely to go extinct at low nutrient concentrations and

333 extreme temperatures in the SRM model than in the RM model. Overall, system persistence for
334 the SRM model was 44% without considering the extinction threshold and 37% when
335 considering it. In other words, without considering extinctions driven by very low biomass
336 densities, the SRM model predicts lower persistence of the consumer compared to RM model but
337 it is the opposite pattern when considering extinctions driven by very low biomass densities. We
338 thus conclude that the RM model predicts larger population fluctuations leading to high
339 probabilities of populations extinctions in comparison to the SRM model.

340

341 *Biomass distribution*

342 We next compared the predictions of both models for consumer-resource biomass ratios along the
343 temperature and nutrient gradients (Fig. 2). We found that the RM model systematically predicts
344 biomass ratio > 1 (i.e. consumer biomass is larger than resource biomass). In contrast, the SRM
345 model predicts biomass ratios both $>$ or $<$ than 1 depending on temperature and nutrient levels.
346 The RM model predicts that, as soon as the consumer can persist, its population biomass density
347 always exceeds the resource population biomass density (Fig. 2). With the SRM model, the
348 biomass ratios are below one at low nutrient levels (Fig. 2). However, at medium and high
349 nutrient levels, the ratios are above one as soon as the consumer can persist. We found
350 qualitatively similar results when considering unstable equilibrium points (Fig. S2). Finally, we
351 showed that, for equivalent parameter values, the RM model predicts biomass ratio that are
352 superior or equal to the ones predicted by the SRM model (text S2). This difference between the
353 two models is independent of the shape and position of the temperature function used to
354 parametrise the models.

355

356 *Mechanisms underlying stability and biomass distribution patterns*

357 Here, we detail the mechanisms underlying the stability and biomass distribution patterns to
358 better understand how and when stoichiometric constraints modulate the effects of temperature
359 and nutrients on consumer-resource dynamics. The first mechanism corresponds to the effect of
360 stoichiometric constraints on the consumer energetic efficiency that determines the consumer
361 persistence at extreme low and high temperatures. The second mechanism relates to the influence
362 of the stoichiometric constraints on population dynamical feedback that explains why the
363 stoichiometric model predicts more stability at high nutrient levels compared to the non-
364 stoichiometric model.

365

366 *Consumer energetic efficiency*

367 The persistence of the consumer at low and high temperatures is driven by the energetic
368 efficiency EE of the consumer (i.e. its feeding rate relative to metabolic losses) calculated as
369 follows:

$$370 \quad EE = \frac{ef(R^*)}{m} \quad (9)$$

371 Where $f(R^*)$ is the functional response of the consumer at resource density R^* (i.e. the resource
372 equilibrium density in absence of the consumer). We recall that the assimilation efficiency e is a
373 function of resource quality Q_R in the SRM model whereas it is constant in the RM model. The
374 intuitive interpretation of eqn. 9 is that EE should be above one for the consumer population to
375 grow and persist.

376

377 To better understand the influence of stoichiometric constraints on consumer persistence, we thus
378 investigated differences in the RM and SRM model predictions regarding the consumer energetic
379 efficiency EE along the temperature gradient at two nutrient concentrations (Fig. 3). For both
380 models, energetic efficiency at equilibrium has a hump-shaped relationship with temperature with

381 maximal efficiency values at medium temperatures. While this unimodal shape is conserved
382 across nutrient levels and models, the RM model systematically predicts higher consumer
383 energetic efficiency values than the SRM model because consumer assimilation efficiency is
384 lower in the SRM than in the RM model (Fig. S3). As a result, the temperatures at which
385 energetic efficiency falls below one and drives consumers extinct are more extreme in the RM
386 model compared to the SRM model (Fig. 3). In other words, energetic efficiency is above one for
387 a narrower thermal range in the SRM model.

388

389 *Dynamical feedbacks due to the stoichiometric constraints*

390 The second mechanism by which stoichiometric constraints influence consumer-resource stability
391 and biomass distribution are the dynamical feedbacks due to stoichiometric constraints on the
392 resource population growth rate and on the consumer energetic efficiency. In the SRM model, the
393 growth rate of the resource population depends on both the total nutrient load and the consumer
394 population density as $Q_R = (N_{\text{tot}} - Q_C C)/R$. In other words, when consumer population increases,
395 this decreases resource population growth by reducing both resource density (through predation)
396 and quality (through the total nutrient load) leading to a negative feedback on consumer
397 population growth rate. In contrast, for the RM model, the negative consumer feedback is only
398 driven by the reduction in resource density as resource quality is not considered. In addition to
399 this first dynamical feedback, there is a second dynamical feedback as the consumer population
400 growth rate also depends on Q_R and thus on its own biomass density. Thus, also this second
401 negative feedback loop limits the consumer population growth rate when its density increases.
402 Altogether, dynamical feedbacks reduce strongly the amplitude of population fluctuations, which
403 in turn increases resource and consumer persistence.

404

405 To reveal the dynamic effects of the stoichiometric constraints, we calculated the values of
406 assimilation efficiencies and carrying capacities predicted by the SRM model for each
407 temperature-nutrient scenario (Fig. S3) and used these effective parameter values to replace the
408 values of parameters e and K in the RM model for each temperature-nutrient scenario. In other
409 words, we calculated average values of e and K in the dynamic SRM model and used them as
410 constant input parameters in the RM model. The objective of using these effective parameter
411 values was to disentangle the static effect of stoichiometric constraints (i.e. changing the average
412 parameter values of consumer assimilation efficiency and of the resource carrying capacity) from
413 their population dynamical effect (i.e. the two dynamical feedback described above). We thus
414 simulated population dynamics along the temperature-nutrient gradient using the RM model with
415 these effective parameters; referred hereafter as effective RM model (Fig. 4). Comparing
416 predictions from the RM, effective RM, and SRM models allowed to disentangle the static
417 stoichiometric effects when going from the RM to the effective RM predictions (Fig. 4, panels a
418 to b) from the dynamical stoichiometric effects when going from the effective RM to the SRM
419 predictions (Fig. 4, panels b to c). In other words, the RM and effective RM only differ in their
420 parameter values because the effective RM takes into account the effect of stoichiometric
421 constraints on the average parameter values. On the other hand, the effective RM and SRM have
422 similar parameter values but different population dynamics, which helps understanding the
423 dynamical feedback induced by stoichiometric constraints.

424
425 We found that, at low nutrient concentrations, population fluctuations and consumer persistence
426 predicted by the effective RM model agreed with predictions of the SRM model. However, the
427 system shifted from a stable equilibrium point to a limit cycle at lower nutrient concentrations for
428 the effective RM model than for the SRM model. This suggests that more nutrients are needed to
429 destabilize the system with the SRM model. Moreover, the effective RM model predicts ampler

430 population fluctuations than the SRM model. As a result, the effective RM predicts high
431 extinction rates at high nutrient concentrations compared to the SRM model. Overall, we found
432 that the effective RM model cannot fully reproduce the dynamics predicted by the SRM, which
433 indicates that including stoichiometric constraints in the RM model involves more than only
434 changing parameter values.

435

436

437

438

439 **Discussion**

440 Temperature and nutrient enrichment are two of the most important drivers of global change
441 (Nelson 2005). However, most research on the effects of temperature and nutrients on community
442 dynamics assumes that the elemental composition of primary producers and consumers are
443 constant and independent of changes on energy and material fluxes (Binzer *et al.* 2012; Boit *et al.*
444 2012; Amarasekare & Coutinho 2014; Gilbert *et al.* 2014; Amarasekare 2015; Binzer *et al.* 2016;
445 Gilarranz *et al.* 2016). Yet, the elemental composition of primary producers is known to be
446 flexible, which can have important consequences for community dynamics and ecosystem
447 processes (Elser *et al.* 2000). We have shown how stoichiometric constraints that account for
448 flexible stoichiometry can affect predictions on how temperature and nutrients influence
449 community stability and biomass distribution across trophic levels. We thus argue that
450 considering stoichiometric constraints is an important step toward a better understanding of the
451 effects of global change on ecosystems.

452

453

454 *Stoichiometric constraints and temperature can dampen the paradox of enrichment*

455 We showed that both stoichiometric constraints and temperature dampen the negative effect of
456 nutrient enrichment on consumer-resource fluctuations and increase system persistence at high
457 nutrient levels. Temperature effects are driven by physiological mechanisms. In agreement with
458 previous empirical studies, our model parametrization reflects the observation that metabolic loss
459 rates increase faster with warming than consumer feeding rates (Vucic-Pestic *et al.* 2011; Sentis
460 *et al.* 2012; Fussmann *et al.* 2014; Iles 2014). Consumers are thereby less energetically efficient
461 at higher temperatures which stabilizes food-web dynamics by reducing energy flow between
462 trophic levels (Binzer *et al.* 2012; Kratina *et al.* 2012; Fussmann *et al.* 2014; Sentis *et al.* 2017).
463 In contrast, the effect of stoichiometric constraints is mainly linked to two mechanisms: a shift in
464 the position of the Hopf bifurcation and negative dynamical feedbacks of the consumer and
465 resource on their population growth rates. Both resources and consumers are composed of the
466 same essential elements (N, P, and C), which implies that, when consumer or resource population
467 biomass increases, it reduces the pool of free nutrients available for the growth of the resource
468 population. Therefore, more nutrients are needed to shift the system from a stable equilibrium to
469 population cycles. In other words, the paradox of enrichment is displaced to higher nutrient
470 concentrations (i.e., the position of the Hopf bifurcation is shifted to higher nutrient levels. In
471 contrast, the RM model does not take into account the storage of nutrients in both the resource
472 and consumer biomasses (i.e. the carrying capacity only depends on the total nutrient load). Less
473 enrichment is thus required to shift the system from a stable equilibrium point to limit cycles.

474
475 We found two dynamic effects that correspond to negative dynamical feedbacks of the consumer
476 and the resource on themselves. When consumer population increases, it decreases the population
477 growth rate of the resource by limiting nutrient availability, diminishing resource biomass, which,
478 in turn, decreases the consumer population growth rate. Conversely, when the resource biomass

479 increases, this decreases the nutrient content of the resource, which, in turn, limits the growth
480 rates of both the resource and consumer populations. These stoichiometric negative feedback
481 loops strongly decrease the amplitude of population fluctuations and thus dampen the paradox of
482 enrichment. Interestingly, our comparisons of the RM, effective RM and SRM model predictions
483 indicate that the dynamical effects contribute more to the reduction of fluctuations than the static
484 effects: population fluctuations are large in the effective RM model accounting for the static
485 effect only, whereas they are much smaller in SRM model accounting for both static and
486 dynamical effects (Fig. 4). This implies that the impact of stoichiometric constraints on
487 community dynamics goes beyond a simple modification of parameter values and encompass
488 more complex population feedbacks between the consumer and the resource.

489
490 Overall, these results demonstrate that both flexible stoichiometry and temperature can
491 synergistically dampen the paradox of enrichment by two different mechanisms: population
492 dynamic feedbacks and physiological constraints. Our consumer-resource model is simplified
493 compared to natural communities composed of numerous species. Moreover, in natural systems,
494 a large amount of nutrient can be stored in abiotic and slow biotic pools that have long turnover
495 times which, in turn, can influence the population dynamics. In particular, the amplitude of the
496 population fluctuations is expected to be smaller as abiotic pools can buffer the population
497 feedback. Nevertheless, considering the nutrient held in slow abiotic or biotic pools would not
498 change the equilibrium densities of primary producers and grazer if nutrients are released in the
499 environment proportionally to their density stored in the abiotic pool. Moreover, the predictions
500 of the stoichiometric model fit with empirical observations. In eutrophic lakes and experimental
501 mesocosms, populations can persist at relatively high nutrient concentrations even if fertilisation
502 enhance population fluctuations (O'Connor *et al.* 2009; Boit *et al.* 2012; Kratina *et al.* 2012), as
503 our stoichiometric model predicts. In contrast, the Rosenzweig-MacArthur model tends to

504 produce very large population fluctuations and extinctions at low nutrient concentrations which
505 can explain why these predictions are not well supported by empirical observations (McAllister *et*
506 *al.* 1972; Jensen & Ginzburg 2005).

507
508 *Effects of stoichiometric constraints on system persistence across environmental gradients*
509 While stoichiometric constraints dampen the paradox of enrichment and thus increase persistence
510 at high nutrient levels, they also reduce the persistence of the consumer at low and high
511 temperatures. Stoichiometric constraints affect the thermal thresholds for consumer extinctions.
512 Consumers can only persist over a narrower range of intermediate temperatures when they are
513 constrained by stoichiometry. This is due to the reduced biomass assimilation of the consumer at
514 low and high temperatures that, in turn, decreases its energetic efficiency and thus fastens
515 consumer extinction. In our stoichiometric model, the reduction of biomass assimilation
516 efficiency emerges from the effect of temperature on resource quality: extreme high and low
517 temperatures decrease resource quality and thus less resource biomass can be converted in
518 consumer biomass at these temperatures. The emergence of a thermal dependency for
519 assimilation efficiency contrasts with previous theoretical studies that used the RM model and
520 assumed that the assimilation efficiency is temperature independent as resource quality is
521 assumed constant (Binzer *et al.* 2012; Gilbert *et al.* 2014; Sentis *et al.* 2017; Uszko *et al.* 2017).
522 In the SRM model, the thermal dependency of the consumer assimilation efficiency is fully
523 driven by the change in the resource stoichiometry induced by temperature. The SRM model thus
524 predicts an additional mechanism by which temperature can influence trophic interactions:
525 temperature changes resource stoichiometry, which, in turn, impacts the consumer assimilation
526 efficiency and its population growth rate. This prediction matches with empirical results showing
527 that primary producer stoichiometric composition can change with temperature (Woods *et al.*
528 2003) and that consumer assimilation efficiency is sensitive to resource stoichiometric

529 composition (Andersen 1997; Elser *et al.* 2000). To sum up, the overall effect of stoichiometric
530 constraints on system persistence thus depends on the temperature range considered and on their
531 relative influence on population fluctuations versus consumer persistence.

532

533 *Effects of stoichiometric constraints on biomass distribution*

534 We found that stoichiometric constraints can modulate the effects of temperature and nutrients on
535 biomass distribution across trophic levels. Without stoichiometric constraints (i.e. with the
536 Rosenzweig-MacArthur model), biomass ratios are above one for almost all temperatures or
537 nutrient levels as the biomass produced by the resource is efficiently transferred to the consumer
538 level consistently along the environmental gradients. This finding agrees with theoretical studies
539 reporting that Lotka-Volterra and RM models predict biomass ratios above one and fail to
540 reproduce biomass pyramids for a substantial region of parameter values (Jonsson 2017; Barbier
541 & Loreau 2019). However, in nature, consumer-resource biomass ratios are often below one
542 (McCauley & Kalff 1981; Del Giorgio & Gasol 1995; McCauley *et al.* 1999; Irigoien *et al.* 2004)
543 suggesting that additional mechanisms should be included to better understand and predict
544 biomass distribution patterns in natural food webs. Our stoichiometric model agrees with
545 experimental observations. It predicts that, at low nutrient concentrations (i.e. $< 0.01 \text{ gP.m}^{-3}$), the
546 biomass ratio never exceeds one along the entire temperature gradient. This is observed in
547 oligotrophic aquatic systems where primary production is too low to sustain high consumer
548 populations (O'Connor *et al.* 2009). In addition, we also found that increasing nutrient levels
549 decreased the temperature ranges within which biomass ratio is below one. This corresponds to
550 results from manipulated nutrient concentrations and temperature in aquatic mesocosms, where
551 zooplankton to phytoplankton biomass ratio only exceeds one in the enriched mesocosms at
552 medium or warm temperatures (i.e. 27°C) (O'Connor *et al.* 2009). This suggests that the models
553 with stoichiometric constraints better reproduce the biomass patterns observed in experimental

554 and natural systems. Nevertheless, further experiments investigating the links between
555 stoichiometric flexibility and consumer-resource dynamics are needed to determine if these
556 stoichiometric mechanisms are underlying patterns of biomass distribution in nature.

557

558 *Implications of our findings for global change*

559 Temperature and nutrients do not act in isolation from each other. Climate warming, for example,
560 causes stronger water stratification which, in turn, can limit nutrient cycling (Sarmiento *et al.*
561 2004; Tranvik *et al.* 2009). Environmental policies such as the European water framework
562 directive (i.e. Directive 2000/60/EC of the European Parliament and of the Council establishing a
563 framework for the Community action in the field of water policy) effectively reduces input of
564 nutrients in aquatic ecosystems (Anneville *et al.* 2005) while the climate keeps warming. With
565 these two phenomena, water will often be warmer and contain fewer nutrients in aquatic systems.
566 Our models consistently predict that warmer temperatures should stabilise consumer-resource
567 dynamics but, if temperature further increases, the consumer goes extinct as energetic efficiency
568 decreases with warming. Moreover, we found that stoichiometric constraints can reduce this
569 thermal extinction threshold (i.e. the consumer persists in a narrower thermal range), especially at
570 low nutrient levels. Our stoichiometric model thus suggests that decreasing nutrient
571 concentrations alongside warmer temperatures should fasten the extinction of consumer
572 populations. This prediction matches empirical observations of consumer extinctions at warm
573 temperatures in oligotrophic aquatic systems (Petchey *et al.* 1999; O'Connor *et al.* 2009).
574 Altogether, these results indicate that considering stoichiometric constraints can be of importance
575 for the management of nutrient inputs and the conservation of natural populations and
576 communities under climate change.

577

578 *Conclusion*

579 Knowledge of how temperature and nutrient simultaneously influence the elemental composition
580 of primary producers and consumers is crucial to better understand and predict the effects of
581 global change on species interactions, community dynamics and fluxes of energy and material
582 within and among ecosystems. Here we showed that stoichiometric constraints dampen the
583 negative effect of enrichment on stability by reducing population fluctuations through population
584 dynamics feedbacks. However, stoichiometric constraints also decrease consumer energetic
585 efficiency, which increases consumer extinction risk at extreme temperatures and low nutrient
586 concentrations. Finally, stoichiometric constraints can reverse biomass distribution across trophic
587 level by modulating consumer efficiency and resource population growth rate along the
588 temperature and nutrient gradients. Our study provides a first step in the exploration of the
589 consequences of stoichiometric constraints and temperature on ecological communities. It
590 suggests that accounting for stoichiometric constraints can strongly influence our understanding
591 of how global change drivers impact important features of ecological communities such as
592 stability and biomass distribution patterns.

593

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597

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- 766
- 767

768 **Figure legends**

769 **Fig. 1.** Population fluctuations (consumer biomass coefficient of variation; panels *a* and *b*) and
770 species persistence (number of species; panels *c* and *d*) across the temperature (*y* axis) and
771 nutrient (*x* axis) gradients as predicted by the Rosenzweig-MacArthur (RM; panels *a* and *c*) and
772 by the Stoichiometric Rosenzweig-MacArthur (SRM; panels *b* and *d*) models. In panels *a* and *b*,
773 the white colour corresponds to the temperature-nutrient scenario for which the consumer has
774 gone extinct whereas the orange to red to dark red represent population fluctuations of increasing
775 amplitude. In panels *c* and *d*, in black: both consumer and resource persist; in red: only the
776 resource persists; in orange: none persists. Resource biomass CV is not shown; it is qualitatively
777 similar to the consumer biomass CV as resource and consumer biomass fluctuation are strongly
778 coupled.

779

780 **Fig. 2.** Consumer-resource biomass ratio along the temperature gradient for the Rosenzweig-
781 MacArthur (RM, green lines) and the Stoichiometric Rosenzweig-MacArthur (SRM, black lines)
782 models at three nutrient concentrations (0.008, 0.02, and 0.032 gP.m⁻³). In each panel, the dotted
783 line represents biomass ratio of one; i.e. the biomass densities of the resource and the consumer
784 are equal. Biomass values shown at equilibrium points. For unstable equilibrium points (i.e. limit
785 cycles), see Fig. S2.

786

787 **Fig. 3.** Consumer energetic efficiency along the temperature gradient for the Rosenzweig-
788 MacArthur (RM, in green) and the Stoichiometric Rosenzweig-MacArthur (SRM, in black)
789 models at two nutrient concentrations (0.008 and 0.02 gP/m³). In each panel, the dotted line
790 represents energetic efficiency equal to one.

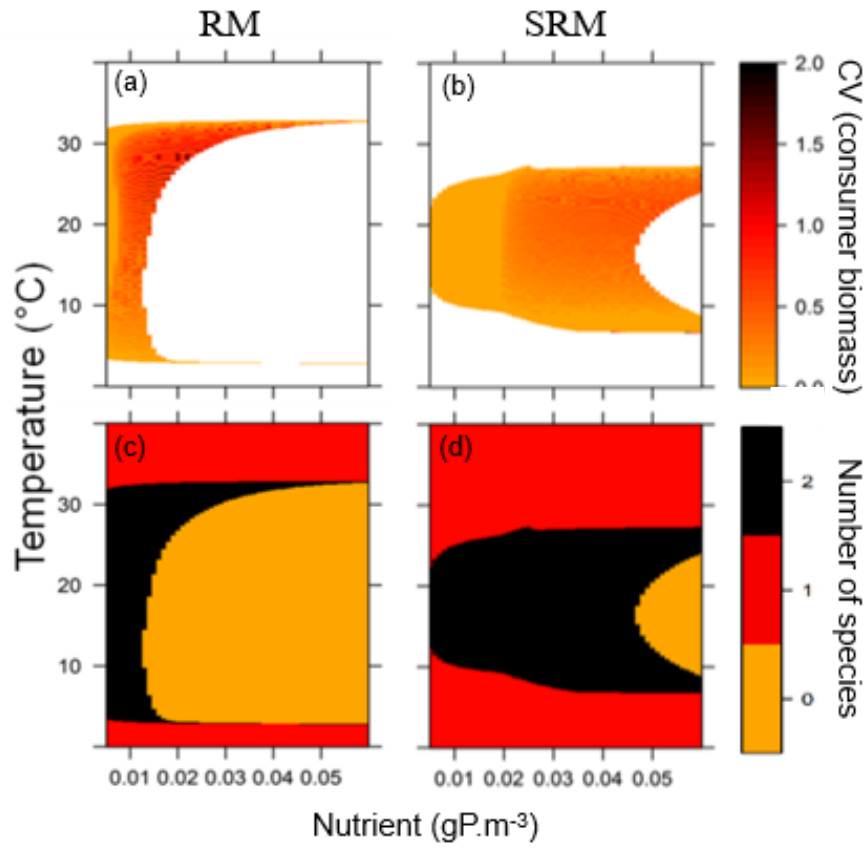
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792 **Fig. 4.** Population fluctuations (consumer biomass coefficient of variation) across the temperature
793 (y axis) and nutrient (*x* axis) gradients as predicted by the Rosenzweig-MacArthur (RM; panel *a*),
794 the RM with effective parameters (panel *b*), and the Stoichiometric Rosenzweig-MacArthur
795 (SRM; panel *c*) models.

796

797 **Figure 1**

798

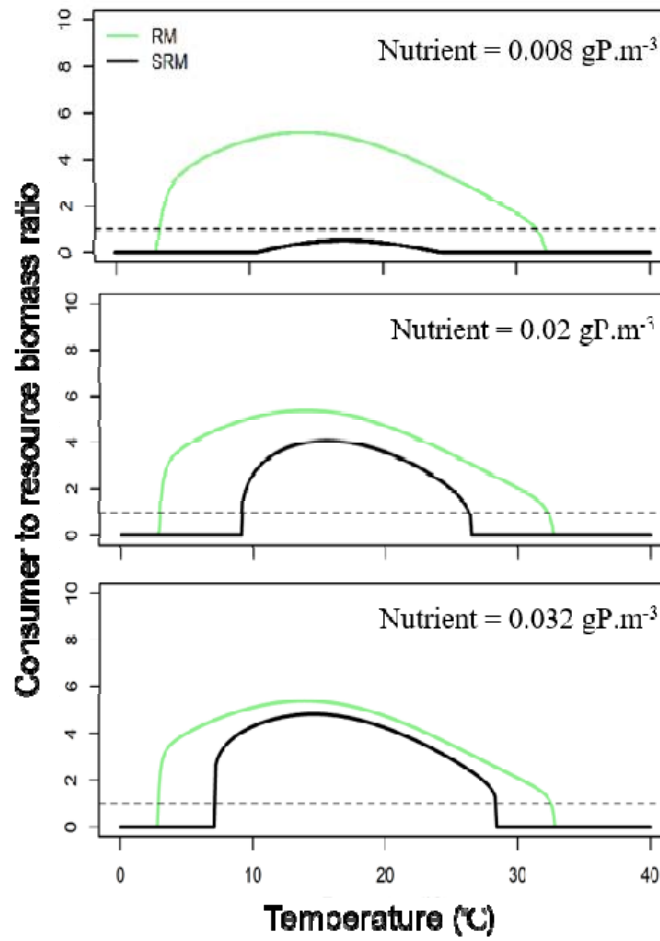


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802 **Figure 2**

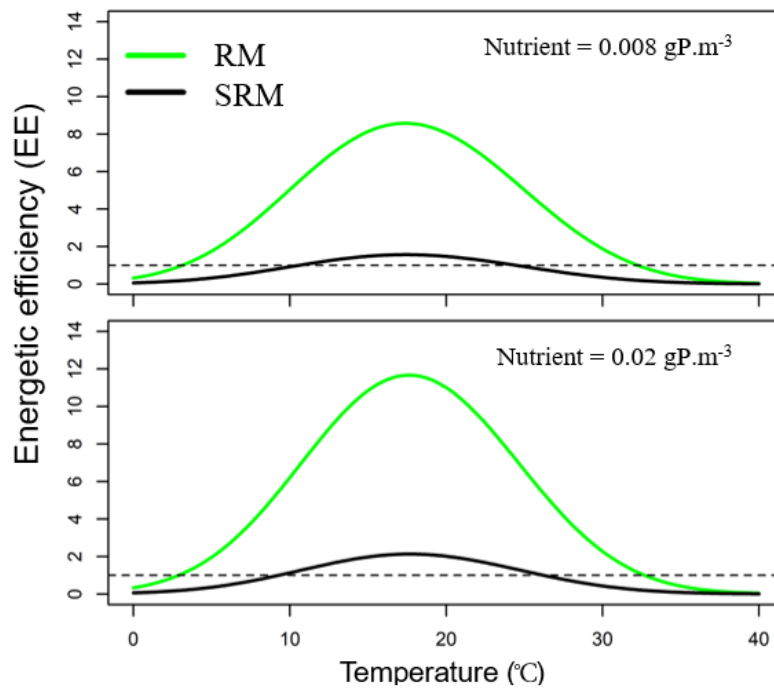


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805 **Figure 3**

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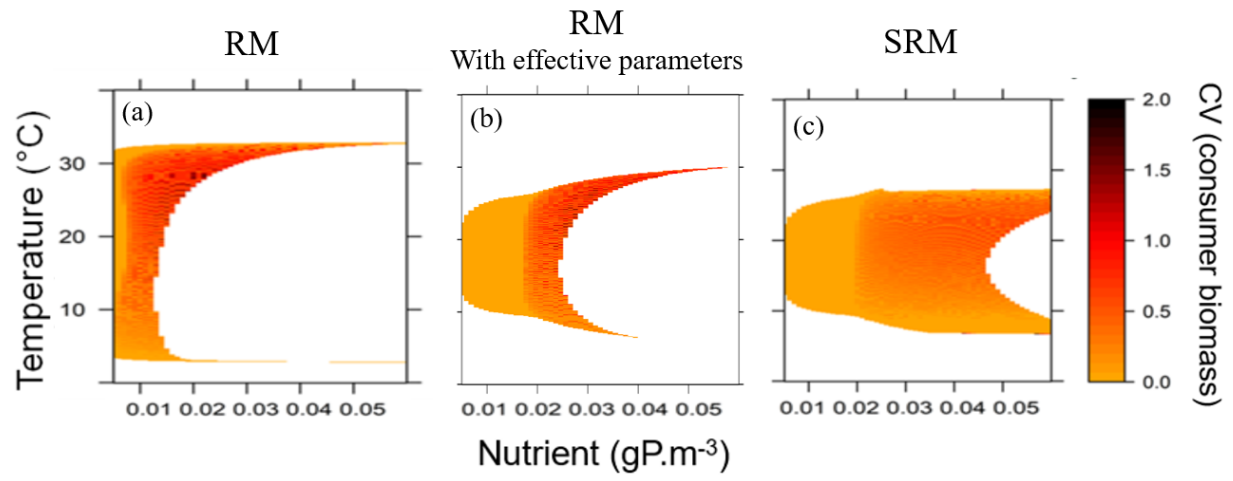
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811 **Figure 4**

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Supplementary information

818 **Stoichiometric constraints modulate the effects of temperature and nutrients on biomass** 819 **distribution and community stability**

820

821

Arnaud Sentis, Bart Haegeman, and José M. Montoya

822

823 **Text S1.** Derivation of the Stoichiometric Rosenzweig-MacArthur (SRM) model

824 The model studied in the main text is very similar to previous stoichiometric consumer-resource
825 models (Andersen 1997; Loladze *et al.* 2000; Andersen *et al.* 2004; Diehl *et al.* 2005). To make
826 our paper self-contained, we here present the model assumptions and derive the model equations
827 (eqs. 3-6 in main text). Our objective was not to develop a complex and very realistic
828 stoichiometric model that would include additional important abiotic and biotic features such as
829 light intensity (Diehl 2007) or compensatory feeding (Cruz-Rivera & Hay 2000). Instead, we
830 aimed at introducing two fundamental stoichiometric features (i.e. stoichiometric flexibility and
831 stoichiometric imbalance) and investigate how these stoichiometric considerations can change
832 predictions of the Rosenzweig-MacArthur model. We assumed that resource and consumer
833 production are limited by energy and a single mineral nutrient. Moreover, we assume the system
834 is closed for nutrients. Thus, nutrient supply originates exclusively from excretion and
835 remineralization of biomass. The total amount of nutrients in the system (N_{tot}) is then a measure
836 of nutrient enrichment. As elemental homeostasis is much stronger for consumers compared to
837 primary producers (Andersen 1997), we assumed the nutrient quota of the consumer Q_C to be
838 constant whereas the nutrient quota of the resource Q_R is flexible. Four differential equations
839 determine the dynamics of four state variables, that is, the concentrations of resource (R) and
840 consumer (C) carbon biomasses and of dissolved mineral nutrients (N), and the nutrient quota of
841 the resource (Q_R):

$$842 \quad \dot{C} = \left(e \frac{aR}{1+ahR} - m \right) C \quad (\text{S1})$$

$$843 \quad \dot{R} = r \left(1 - \frac{Q^{\text{min}}}{Q_R} \right) R - \frac{aR}{1+ahR} C \quad (\text{S2})$$

844 $\dot{Q}_R = h(N) - r(Q_R - Q_{min})$ (S3)

845 $\dot{N} = (Q_R - Q_C e) \frac{aR}{1+ahR} C + Q_C m C - h(N)R$ (S4)

846 As in the RM model, rates of change of the consumer and resource biomass densities \dot{C} and \dot{R}
847 depend on their respective carbon biomass densities C and R ($\text{gC}\cdot\text{m}^{-3}$), except that the resource
848 population growth rate follows the Droop equation (Droop 1974) and is now limited by its
849 nutrient quota Q_R relative to the minimum nutrient quota Q_{min} . Rate of change of Q_R depends on
850 the nutrient uptake rate by the resource species $h(N)$ and the amount of nutrient invested in
851 growth (eqn S3). $h(N)$ is the specific resource nutrient uptake rate and can be represented by a
852 Michaelis-Menten model where the amount of nutrient uptake saturates at high nutrient
853 concentrations.

854
855 With the mass-balance equation, we get that the total amount of nutrient is the sum of the free
856 nutrient plus the nutrient fixed in the resource biomass plus the nutrient fixed in the consumer
857 biomass: $N_{tot} = N + Q_R R + Q_C C$. As Eqns S1-S4 conserve total biomass (the system is closed), the
858 time derivative of N_{tot} is zero. We can thus replace one of the four differential equations S1-S4
859 with the algebraic equation $N_{tot} = N + Q_R R + Q_C C$:

860 $\dot{C} = (e \frac{aR}{1+ahR} - m)C$

861 (S5)

862 $\dot{R} = r(1 - \frac{Q_{min}}{Q_R})R - \frac{aR}{1+ahR}C$

863 (S6)

864 $\dot{N} = (Q_R - Q_C e) \frac{aR}{1+ahR} C + Q_C m C - h(N)R$

865 (S7)

866 $N_{tot} = N + Q_R R + Q_C C$

867 (S8)

868

869 It is possible to derive a simpler model by reducing the number of dimensions in the above model
 870 from three to two. This model reduction is based on the assumption that free nutrients are taken
 871 up very quickly relative to the dynamics of the consumer and resource biomasses. This
 872 corresponds to taking $h(N)$ large, $h(N) = \frac{1}{\varepsilon} \tilde{h}(N)$ for small ε . The fast dynamics (on the
 873 timescale $t \sim \varepsilon$) are

$$874 \quad \dot{N} = -\frac{1}{\varepsilon} \tilde{h}(N)R + (\text{slower contributions})$$

875 (S9)

876 Which converge to $N \rightarrow 0$, and $Q_R \rightarrow \frac{N_{tot} - Q_C C}{R}$ with N_{tot} the total nutrient in the system. In other
 877 words, N in dead and excreted matter is immediately recycled and acquired by the resource
 878 species. When substituting the quasi-steady-state in eqns. (S5, S6), we get the resulting dynamics
 879 (on the timescale $t \sim 1$):

$$880 \quad \dot{C} = \left(e \frac{aR}{1+ahR} - m \right) C \tag{S10}$$

$$881 \quad \dot{R} = r \left(1 - \frac{Q_{min}}{Q_R} \right) R - \frac{aR}{1+ahR} C \tag{S11}$$

$$882 \quad N_{tot} = Q_R R + Q_C C \tag{S12}$$

883 From the nutrient conservation equation (eqn. S12) we obtain that $Q_R = \frac{N_{tot} - Q_C C}{R}$. The intuitive
 884 interpretation is that the resource nutrient quota Q_R decreases with the density of the resource
 885 population and with the density of nutrient stored in the consumer biomass. In contrast to eqns
 886 S5-S8, the reduced model has only two differential equations and one algebraic equation. It can
 887 be equivalently written as a set of three differential equations with \dot{C} and \dot{R} similar as equations

$$888 \quad \text{S10 and S11 and with } \dot{Q}_R = \frac{d \frac{N_{tot} - Q_C C}{R}}{dt} = (Q_R - Q_C e) \frac{aR}{1+ahR} C + \frac{Q_C m C}{R} - r(Q_R - Q_{min}).$$

889

890 In the RM model, the growth rate of the consumer population is assumed to depend only on
891 resource density. We relaxed this assumption by making the population growth rate of the
892 consumer dependent on both the resource quality (i.e. nutrient quota) and quantity (i.e. density).
893 In the SRM model, consumer production is also limited by resource quality as the consumer
894 assimilation efficiency e is a saturating function of resource nutrient quota Q_R :

$$895 \quad e(Q_R) = e_{max} \frac{Q_R}{Q_R + Q_C} \quad (S13)$$

896 The intuitive interpretation of eqn. S13 is that resource quality is not a limiting factor for
897 consumer growth as long as the nutrient content of the resource is superior to the nutrient content
898 of the consumer (i.e. $Q_R > Q_C$). In other words, when $Q_R \gg Q_C$, $e(Q_R) \rightarrow e_{max}$ and when $Q_R \ll$
899 Q_C , $e(Q_R) \rightarrow 0$. By replacing e by $e(Q_R)$ in eqn. S10, we obtain the SRM model.

900

901 **Text S2. Differences in biomass ratios predicted by the two models**

902 Here we show that the equilibrium consumer-to-resource biomass ratio in the model with
 903 stoichiometric constraints (SRM model) is always smaller than the one in the model without
 904 stoichiometric constraints (RM model), keeping the same parameter values. For simplicity we
 905 assume for both models that the consumer and the resource persist at equilibrium, and we do not
 906 consider the stability of the equilibrium point (in particular, the equilibrium might be unstable at
 907 the center of a limit cycle). We indicate the equilibrium values of the non-stoichiometric model
 908 by the superscript “ns” and the equilibrium values of the stoichiometric model by the superscript
 909 “s”. We use the same superscripts to distinguish the assimilation efficiencies of both models.

910 *Model without stoichiometric constraints*

911 The model is defined as

$$912 \quad \dot{R} = r\left(1 - \frac{R}{K}\right)R - f(R)C \quad (\text{S14})$$

$$913 \quad \dot{C} = (e^{ns}f(R) - m)C \quad (\text{S15})$$

914

915 With $f(R) = \frac{aR}{1+ahR}$ and $K = N_{tot}/Q_{min}$.

916 By solving equation (S15) we get the resource biomass at equilibrium:

$$917 \quad R^{ns} = \frac{1}{a\left(\frac{e^{ns}}{m} - h\right)} \quad (\text{S16})$$

918 From equation (S14) we get the consumer biomass at equilibrium. It follows from $f(R^{ns})C^{ns} =$
 919 $r\left(1 - \frac{R^{ns}}{K}\right)R^{ns}$, or $C^{ns} = r\left(1 - \frac{R^{ns}}{K}\right)\frac{R^{ns}}{f(R^{ns})}$

920 Hence, the consumer-to-resource biomass ratio is

$$921 \quad \frac{C^{ns}}{R^{ns}} = \frac{re^{ns}}{m}\left(1 - \frac{R^{ns}}{K}\right) = \frac{re^{ns}}{m}\left(1 - \frac{Q_{min}R^{ns}}{N_{tot}}\right) \quad (\text{S17})$$

922

923 *Model with stoichiometric constraints*

924 The model is defined as

$$925 \quad \dot{R} = r\left(1 - \frac{Q_{min}}{Q_R}\right)R - f(R)C \quad (\text{S18})$$

$$926 \quad \dot{C} = (e^s(Q_R)f(R) - m)C \quad (\text{S19})$$

927

928 With $e^s(Q_R) = e_{max} \frac{Q_R}{Q_C + Q_R}$, $f(R) = \frac{aR}{1+ahR}$ and $N_{tot} = Q_R R + Q_C C$.

929 From equation (S19) we have $e^s(Q_R^s) f(R^s) = m$, or

$$930 \quad R^s = \frac{1}{a\left(\frac{e^s(Q_R^s)}{m} - h\right)} \quad (\text{S20})$$

931 From equation (S18) we have from $(R^s)C^s = r\left(1 - \frac{Q_{min}}{Q_R^s}\right)R^s$, or

$$932 \quad C^s = \frac{re^s(Q_R^s)}{m} \left(1 - \frac{Q_{min}}{Q_R^s}\right) R^s$$

933 Hence, the consumer-to-resource biomass ratio is

$$934 \quad \frac{C^s}{R^s} = \frac{re^s(Q_R^s)}{m} \left(1 - \frac{Q_{min}}{Q_R^s}\right) = \frac{re^s(Q_R^s)}{m} \left(1 - \frac{R^s Q_{min}}{N_{tot} - Q_C C^s}\right) \quad (\text{S21})$$

935

936

937 We now compare the biomass ratios of equations (S17) and (S21). We have $e^{ns} = e_{max}$, and hence

$$938 \quad e^{ns} \geq e^s(Q_R^s). \quad (\text{S22})$$

939 Using this inequality, we get $a\left(\frac{e^{ns}}{m} - h\right) \geq a\left(\frac{e^s(Q_R^s)}{m} - h\right)$, and by equations (S16) and (S20), we

940 see that

$$941 \quad R^{ns} \leq R^s \quad (\text{S23})$$

942 Clearly, we always have $N_{tot} \geq N_{tot} - Q_C C^s$. Combining this with equation (S23), we get

$$943 \quad \frac{R^{ns}}{N_{tot}} \leq \frac{R^s}{N_{tot} - Q_C C^s} \text{ and}$$

$$944 \quad 1 - \frac{Q_{min} R^{ns}}{N_{tot}} \geq 1 - \frac{R^s Q_{min}}{N_{tot} - Q_C C^s} \quad (\text{S24})$$

945 Finally, from equations (S22) and (S24),

$$946 \quad \frac{re^{ns}}{m} \left(1 - \frac{Q_{min} R^{ns}}{N_{tot}}\right) \geq \frac{re^s(Q_R^s)}{m} \left(1 - \frac{R^s Q_{min}}{N_{tot} - Q_C C^s}\right), \quad (\text{S25})$$

947 showing that $\frac{C^{ns}}{R^{ns}} \geq \frac{C^s}{R^s}$.

948 **Table S1.** Definitions and units of model parameters, from Uszko *et al.* (2017). For temperature-
 949 dependent parameters, we list the value of the scaling constant Q_0 (in units of the parameter) and
 950 the values of either the activation energy E_Q (eV, when temperature dependence is monotonous,
 951 eqn. 7) or of the temperature T_{opt} (Kelvin) at which the parameter value reaches a
 952 maximum/minimum and the width s (Kelvin) of this bell-/U-shaped function (when temperature
 953 dependence is non-monotonous, eqn. 8). Biomass and nutrients are expressed in units of carbon
 954 (C) and phosphorus (P), respectively

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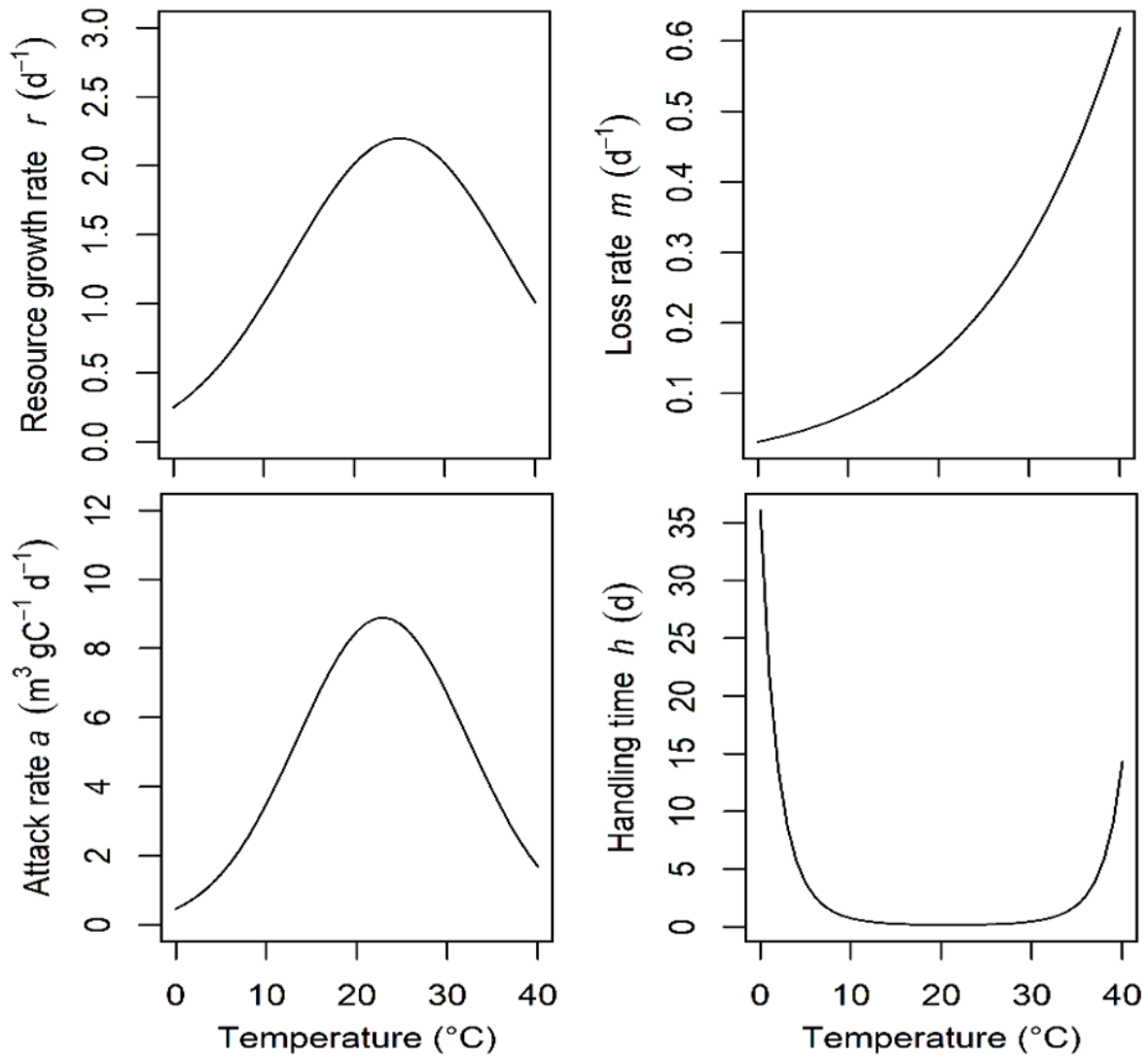
Temperature independent				
Parameter	value	Unit	Definition	Reference
Thermal parameters				
r	$r_0 = 2.2; T_{opt} = 298.15; s = 12.0$	1/d	Intrinsic rate of resource net production (gross production – biosynthesis costs)	Uszko et al. 2017
h	$h_0 = 0.17; T_{opt} = 294.1; s = 6.4$	d	Handling time	Uszko et al. 2017
a	$a_0 = 8.9; T_{opt} = 296.0; s = 9.4$	$m^3/(gC\ d)$	Attack rate	Uszko et al. 2017
m	$m_0 = 4.4 \times 10^8; E_m = 0.55$	1/d	Consumer mortality plus maintenance rate	
e_{max}	0.385	-	Maximum assimilation efficiency	Yodzis
Q_C	0.042	g P/g C	Consumer P:C ratio	Diehl 2005
Q_{min}	0.009	g P/g C	Minimum nutrient quota	Diehl 2005
Q_R	Variable	g P/g C	Resource P:C ratio	
N_{tot}	Variable	g P/m ³	Total nutrients in the system	
T	Variable	K	Temperature	

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959 **Fig. S1.** Thermal functions used to parametrize the model (adapted from Uszko et al. 2017)



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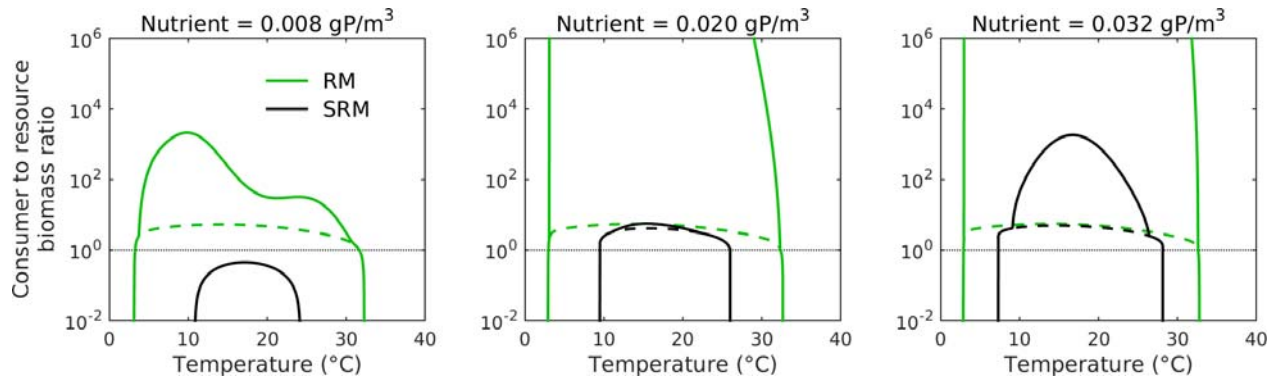
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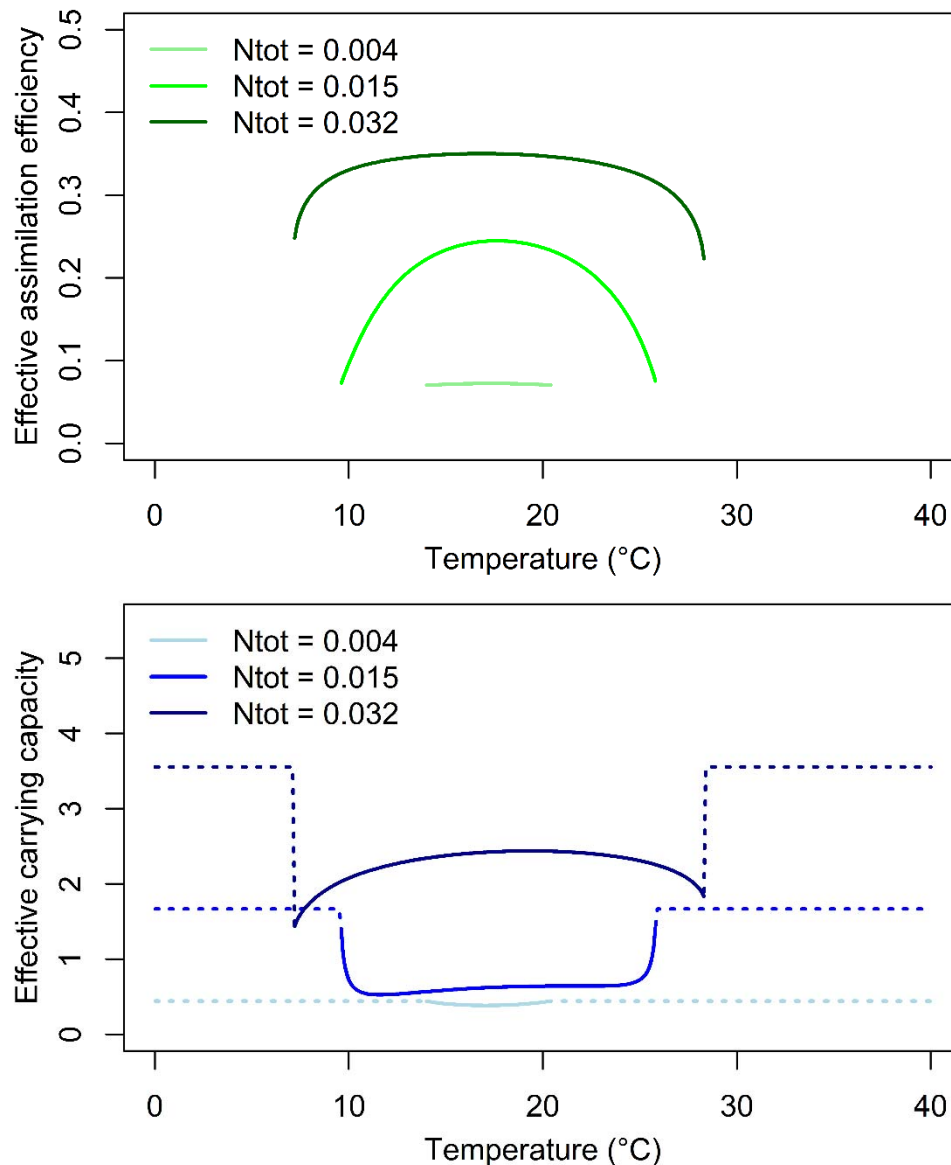
968 **Fig. S2.** Consumer-resource biomass ratio (log scale) along the temperature gradient for the
969 Rosenzweig-MacArthur (RM, green lines) and the Stoichiometric Rosenzweig-MacArthur (SRM,
970 black lines) models at three nutrient concentrations (0.008, 0.02, and 0.032 $\text{gP}\cdot\text{m}^{-3}$). In each
971 panel, the dotted lines represent unstable solutions whereas full lines represent stable solutions.
972 The thin horizontal dotted line represents biomass ratio of one; i.e. the biomass densities of the
973 resource and the consumer are equal.

974

975



977 **Fig. S3.** Effective assimilation efficiency e_{ef} and carrying capacity K_{ef} from the
978 Stoichiometric Rosenzweig-MacArthur (SRM) model along the temperature gradient at
979 three nutrient levels (0.004, 0.015, and 0.032 gP.m⁻³) with $Q_C = 0.042$. Full lines represent
980 temperature and nutrient scenarios for which both the resource and consumer persist
981 whereas dotted lines represent scenarios for which only the resource persists. Effective
982 assimilation efficiency was calculated as $e_{ef} = e_{max}Q_R/(Q_R+Q_C)$, with Q_R the equilibrium
983 solution of the SRM model and the effective carrying capacity as $K_{ef} = Q_R R/Q_{min} = (N_{tot} -$
984 $Q_C C)/Q_{min}$, with Q_R , R and C the equilibrium solutions of the SRM model.



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